Mathematical Modelling of Transport Phenomena in Concrete Porous Media

Ilija Plecas

Institute of Nuclear Sciences "Vinca" P.O.BOX 522, 11001 Belgrade, Serbia

E-mail: iplecas@vinca.rs

ABSTRACT

Large parts of repositories for low and intermediate level radioactive waste consist of cementitious materials. Radionuclides are transported by diffusion in the cement matrix or, in case of fractured or highly permeable cement, by advection and dispersion. Two fundamental concerns must be addressed when attempting to isolate low-level waste in a disposal facility on land. The first concern is isolating the waste from water, or hydrologic isolation. The second is preventing movement of the radionuclides out of the disposal facility, or radionuclide migration. Particularly, we have investigated here the latter modified scenario. To assess the safety for disposal of radioactive waste-concrete composition, the leakage of ^{137}Cs from a waste composite into a surrounding fluid has been studied. Leakage tests were carried out by original method, developed in Vinca Institute [1,2,3,4,5]. Transport phenomena involved in the leaching of a radioactive material from a cement composite matrix are investigated using three methods based on theoretical equations. The results presented in this paper are from a 30-year mortar and concrete testing project that will influence the design choices for radioactive waste packaging for a future Serbian radioactive waste disposal center.

Keywords : radioactive waste, cement, concrete, transport phenomena, leakage rate

1 INTRODUCTION

Cement and concrete are widely used in low-level waste management both as a means of solidifying waste and for containment of dry or liquid wastes. At present there is also widespread interest in the use of near-surface concrete trench system for the disposal of radwaste materials. Typical concrete is a mixture of cement, sand, stone aggregate and water in various proportions that together determine the structural properties and tightness of the poured material. Water content is one of the critical parameters and must be carefully controlled during curing and setting; to a large extent it will determine the porosity of the resulting material.

Engineered barriers are features of the disposal system made or altered by humans during the construction, operation and closure of a repository. Engineered barriers are intended to contribute to the overall performance of the disposal system by providing the level of containment required while the waste remains hazardous. In cases where the selected site or geological environment is not ideally suited for disposal, the repository can be heavily engineered so that, for meeting safety targets, reliance is placed primarily on the engineered barriers.

Engineering near-surface disposal system provides three biological protection barriers:

1. Mortar for immobilizing the waste and filling the concrete containers. (enable its

penetration into all cavities of solid radioactive waste and thus fix it permanently; permit no leakage of radionuclides- use of special cement or adsorbers and provide primary biological protection)

2. Concrete container (store and enable transport of low and intermediate level radioactive waste; provide secondary biological protection; provide safe keeping of radioactive waste for 300-500 years)
3. Concrete for filling trenches (provide tertiary biological protection and prevent leakage of radionuclides which have penetrated the second barrier)

At the "Vinca" Institute of Nuclear Sciences, a promising composite for mortar, concrete for concrete container and concrete for engineer trenches system, especially for concrete for filling trenches has been developed. Leakage test for leakage measurements of radionuclides from concrete was studied using the above method [1, 3, 6, 7, 8, 9, 10, 11]

2. THEORETICAL METHODS

Transport phenomena involved in the leaching of a radioactive material from a cement composite matrix are investigated using three methods based on theoretical equations. These are: Method I- the diffusion equation for a plane source model, Method II- rate equation for diffusion coupled with a first-order equation, and an empirical method employing a polynomial equation (Method III).

Three methods are compared with respect to their applicability to experimental leaching data [6, 10] . Transport phenomena involved in the leaching of a radioactive material from a composite matrix into surrounding water were investigated using three methods based on theoretical equations.

2.1. Method I: Diffusion equation based on a plane source model

In this model the fraction f leached at time t (d) :

$$f = \frac{\sum a_n}{A_o} = \frac{2 S \sqrt{D_e t_n}}{V \sqrt{\pi}} \quad (1)$$

Where $\sum a_n$ is the cumulative fraction leached of contaminant for each leaching period, A_o is the initial amount of contaminant in the sample, V is the volume of sample (cm^3), S is the exposed surface area of the sample (cm^2) , t_n the duration of leachant renewal period (d) and D_e is the diffusion coefficient ($\text{cm}^2 \text{d}^{-1}$)

The results may also be expressed by the cumulative fraction of the contaminant. Leach test results are plotted as the cumulative fraction of contaminant leached from the samples as a function of the square root of total leaching time:

$$\frac{\sum a_n}{A_o} \text{ versus } \sqrt{\sum t_n} \quad (2)$$

If the model is applicable a plot of $\sum a_n/A_o$ versus $\sqrt{\sum t_n}$ is a straight line and the diffusion coefficient D_e is given by:

$$D_e = \frac{\pi}{4} m^2 \frac{V^2}{S^2} \quad (3)$$

where

$m = (\sum a_n/A_o) (1/\sqrt{\sum t_n})$, is the slope of the straight line ($\text{d}^{-1/2}$).

2.2. Method II: Rate equation for coupled diffusion and simultaneous first-order Reaction

Equations are derived for the rate of absorption and the concentration when a solute of limited solubility diffuses into a semi-infinite medium with which it undergoes a 1st-order reaction. Consideration is given to the case where the medium is a mixture only 1 component of which reacts.

The model incorporates the effects of dispersion, water flow and retardation. In this model, the rate equation is:

$$\frac{\partial C}{\partial t} = D_e (\partial^2 C / \partial X^2) + g(C) \quad (4)$$

C is the ratio of radionuclide concentration in water to source concentration at some time (t)

D is dispersion coefficient of radionuclide in porous medium (cm² d⁻¹)

X is thickness of granular medium (cm)

Here, the special case where g(C) is directly proportional to the concentration C, i.e. a first-order reaction was considered. The initial and boundary conditions are,

$$t = 0, \quad x > 0, \quad C = C_o \quad (5)$$

$$t = 0, \quad x < 0, \quad C = 0 \quad (6)$$

$$t > 0, \quad x = 0, \quad C = 0 \quad (7)$$

From this, the fraction leached from a specimen having a surface area S(cm²) and volume V(cm³) is

$$f = (S/V) \sqrt{D_e/k} [(kt + 1/2) \operatorname{erf} \sqrt{kt} + \sqrt{kt/\pi} \exp(-kt)] \quad (8)$$

Where k is the rate constant of the first-order reaction and

$$\operatorname{erf}(u) = (2/\sqrt{\pi}) \int_0^u \exp(-z^2) dz \quad (9)$$

2.3. Method III: Polynomial equation

The orthogonal polynomial is one of the most useful empirical equations. Its general form is:

$$y(x) = \sum_{i=1}^n A_i \delta_i(x) \quad (10)$$

where:

A_i- is the parameter to be determined, and

δ_i- is a function of x. Here, φ_i(x) - is taken as t^{i/2}, and the leaching fraction is given by

$$f = \sum_{i=1}^n A_i t^{i/2} \quad (11)$$

To simplify the mathematical treatment, a fourth terms polynomial of the form

$$f = A_0 + A_1 t^{1/2} + A_2 t + A_3 t^{3/2} \quad (12)$$

was fitted to the leaching data. For this type of model, extrapolation to longer term leaching is not advisable since the arbitrary constants do not necessarily have any physical significance.

3 PREPARATION OF SAMPLES FOR LEACHING TEST

More than 100 different formulations of concrete formulations were examined to optimize their mechanical and sorption properties. The concrete samples were prepared with a standard Portland cement PC-20-Z-45 MPa. The concrete composition was mixed with artificial radioactivity of ^{137}Cs , $A_0=55-67(\text{kBq})$. Mixing time was about ten minutes. The mixtures were cast into 50 mm diameter cylindrical molds with a height of 50 mm, which were then sealed and cured for 28 days prior to the leaching experiments. Leaching of ^{137}Cs was studied using the method recommended by the IAEA [12] The duration of leachant renewal period was 60 days. After each leaching period the radioactivity in the leachant was measured using EG&G- ORTEC spectrometry system and software. In this paper we discuss four representative formulations of concrete for filling trenches : C_1 , C_2 , C_3 and C_4 . Composition and initial radioactivity $A_0(B_q)$ per sample prepared for leakage test, is shown in Table 1. The cement, sand and added water make up a "paste", the volume of which should be slightly more than the void volume of the coarse aggregate. Typically, the volume of coarse aggregate (including its void volume) represents about 70-80% of the volume of concrete finally produced.

Table 1. Concrete compositions (calculated as grams for 1000 cm^3 of samples)

Materials	Formula			
	T_1	T_2	T_3	T_4
Cement (Portland),gr	400	410	420	430
Sand, 0-2 mm	850	590	720	660
Aggregate,2-4mm	518	516	500	316
Aggregate,4-8mm	430	680	592	886
Water,ml	180	180	180	180
Additives,ml	4	4	4	4
Initial activity $A_0(\text{kBq})$ ^{137}Cs	Per sample ~60,0			

4 RESULTS

Experimental data show the fractions of ^{137}Cs leached from concrete composite as a function of the square root of the leaching period. The linear relation between f and t is not observed throughout the test period. From the application of Method I to the leaching data we obtained:

$$\begin{aligned}
 f_I(T_1) &= 2,2 \cdot 10^{-5} t^{1/2} + 7,6 \cdot 10^{-9} \\
 f_I(T_2) &= 3,9 \cdot 10^{-5} t^{1/2} + 5,8 \cdot 10^{-9} \\
 f_I(T_3) &= 4,4 \cdot 10^{-5} t^{1/2} + 6,3 \cdot 10^{-9} \\
 f_I(T_4) &= 5,5 \cdot 10^{-5} t^{1/2} + 7,5 \cdot 10^{-9}
 \end{aligned}$$

The diffusion coefficients predicted by Method I are:

$$\begin{aligned}
 D_I(T_1) &= 5,5 \cdot 10^{-6} \text{ (cm}^2\text{/d)} \\
 D_I(T_2) &= 5,9 \cdot 10^{-6} \text{ (cm}^2\text{/d)} \\
 D_I(T_3) &= 6,3 \cdot 10^{-6} \text{ (cm}^2\text{/d)} \\
 D_I(T_4) &= 7,8 \cdot 10^{-6} \text{ (cm}^2\text{/d)}
 \end{aligned}$$

Method II was applied to the leakage data to obtain the unknown parameters D_e and k . From this we obtained:

$$\begin{aligned}
 D_{II}(T_1) &= 3,50 \cdot 10^{-6} \text{ (cm}^2\text{/d)}, k = 4,2 \cdot 10^{-4} \text{ (day}^{-1}\text{)} \\
 D_{II}(T_2) &= 4,90 \cdot 10^{-6} \text{ (cm}^2\text{/d)}, k = 3,2 \cdot 10^{-4} \text{ (day}^{-1}\text{)} \\
 D_{II}(T_3) &= 5,73 \cdot 10^{-6} \text{ (cm}^2\text{/d)}, k = 2,8 \cdot 10^{-4} \text{ (day}^{-1}\text{)} \\
 D_{II}(T_4) &= 6,70 \cdot 10^{-6} \text{ (cm}^2\text{/d)}, k = 2,5 \cdot 10^{-4} \text{ (day}^{-1}\text{)}
 \end{aligned}$$

Using the least squares procedure, Method III yielded:

$$\begin{aligned}
 f_{III}(T_1) &= 7,6 \cdot 10^{-8} + 5,7 \cdot 10^{-5} t^{1/2} + 5,3 \cdot 10^{-8} t + 9,5 \cdot 10^{-12} t^{3/2} \\
 f_{III}(T_2) &= 5,8 \cdot 10^{-8} + 6,8 \cdot 10^{-5} t^{1/2} + 6,6 \cdot 10^{-8} t + 9,4 \cdot 10^{-12} t^{3/2} \\
 f_{III}(T_3) &= 4,3 \cdot 10^{-8} + 7,4 \cdot 10^{-5} t^{1/2} + 6,6 \cdot 10^{-8} t + 9,7 \cdot 10^{-12} t^{3/2} \\
 f_{III}(T_4) &= 3,6 \cdot 10^{-8} + 7,2 \cdot 10^{-5} t^{1/2} + 9,8 \cdot 10^{-8} t + 9,7 \cdot 10^{-12} t^{3/2}
 \end{aligned}$$

Fig.1 and Fig. 2 present plots of f against t for leakage of ^{137}Cs from the four concrete samples, for Methods I and Method III.

Fig.1. Plot of f against t (d) for leakage rate of ^{137}Cs from concrete (Method I) Fig.2. Plot of f against t (d) for leakage rate of ^{137}Cs from concrete (Method III)

5 DISCUSSION AND CONCLUSION

Results are presented in Fig 1. and 2. which shows the fraction of ^{137}Cs leached from cement composites as a function of the square root of leaching period. In the data for cement composite as a matrix, linearity between f and t (d) is not observed throughout the time tested; however, there are two different linearity before and after a leaching time of about 10 days. The slope of the linear relation for the early stage is larger than for the latter one. This change in the leaching rate may be associated with the fact that, as the leaching time elapsed, the diffusion rate would gradually slow down as the diffusion path becomes longer. Method I provides a fairly good approximation at longer leaching period. Method I cannot describe the whole leaching process; but it is very convenient to simulate leaching over a long period because of its simplicity. Despite the very complex numerical treatment required, the fit obtained using Method II is no better than that obtained using Method I. Method III provides the best approximation over the whole leaching period. In many cases, however, the leaching mechanisms are unknown and, therefore, it is convenient to use polynomial approximation. Finally, the results presented in this paper will define the design of our future engineered trenches disposal system for radioactive waste.

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